- 1 Glyphosate-based herbicide (GBH) challenged thermoregulation in lizards (*Eremias argus*),
- 2 but warming could compensate for it.
- 3 Simin Yu^a, Yufan Nie^a, Zikang Wang^a, Luyao Zhang^b, Rui Liu^a, Yuping Liu^a, Hongjun Zhang^c,
- 4 Wentao Zhu^a, Mingqi Zheng^{a*}, Jinling Diao^{a*}
- 5 ^aDepartment of Applied Chemistry, China Agricultural University, Yuanmingyuan West Road 2,
- 6 Beijing 100193, China
- 7 ^b School of Food and Biological Engineering, Shaanxi University of Science and Technology, Xi'an
- 8 710021, China
- 9 ^cInstitute for the Control of Agrochemicals, Ministry of Agriculture and Rural Affairs (ICAMA),
- 10 Beijing 100125, P.R.China
- 11 *Corresponding authors:
- 12 Mingqi Zheng, Department of Applied Chemistry, China Agricultural University, Yuanmingyuan
- 13 west road 2, Beijing 100193, P.R. China; E-mail: mqzheng@cau.edu.cn
- 14 Jinling Diao, Department of Applied Chemistry, China Agricultural University, Yuanmingyuan
- 15 west road 2, Beijing 100193, P.R. China; E-mail: lingyinzi1201@gmail.com
- 16

17 Abstract

The impacts of global climate change and its associated stressors on biodiversity are of mounting 18 19 concern. Glyphosate have attracted worldwide attention due to their ubiquitous occurrence, yet their 20 impact on reptiles remains unknown. Glyphosate-based herbicide (GBH) exposure in reptiles is 21 currently hampering efforts to predict how individual reptiles can buffer climate change. We 22 designed a simulated environmental exposure experiment over 60 days on the lizard (Eremias 23 argus). Preferred body temperature and active body temperature data were collected to calculate the 24 accuracy of thermoregulation, while liver detoxification metabolic enzymes, oxidative stress system 25 function, and the non-targeted metabolome of the brain tissue were assessed. Warmer-treated lizards adjusted their physiological levels and behavioral strategies in response to increased ambient 26 27 temperatures and maintained body temperature homeostasis at moderate thermal perturbations. 28 GBH-treated lizards suffered from oxidative damage to the brain tissue and abnormal histidine 29 metabolism, thus their thermoregulatory accuracy reduced. Interestingly, at elevated ambient 30 temperatures, GBH treatment did not affect on their thermoregulatory, possibly through several 31 temperature-dependent detoxification mechanisms. Importantly, this data suggested that the subtle 32 toxicological effects of GBH may threaten increasingly thermoregulation behavior of E. argus with 33 species-wide repercussions, as climate change and exposure time extension.

- Keywords: herbicide, global climate change, reptile, thermoregulation, temperature-dependent
 toxicity, plastic response
- 36

37 **1.Introduction**

38 Human activities have a profound impact on the natural world, from climate change and habitat destruction to overfishing and the introduction of invasive species. According to reports from the 39 Intergovernmental Panel on Climate Change (IPCC), global warming is no longer in doubt and is 40 41 characterized by a rise in global average temperatures, with temperatures projected to rise by 2.6-42 4.8°C by 2100 (RCP4.5 scenarios)(Stocker and D., 2013). Unfortunately, a large number of species live in highly fragile environments, and many are reaching the limits of their ability to adapt to the 43 44 changing climate, which leads to local and global extinctions and a serious threat to biodiversity 45 (Sinervo et al., 2010a). 46 As ectotherms, reptiles are strongly affected by global warming, where environmental conditions 47 strongly influence their life history through critical behaviors and physiological processes. As there 48 is an optimum temperature range for physiological and biochemical reactions, too high or too low 49 body temperatures can disrupt an individual's metabolism and ultimately lead to death. For whole 50 populations or species, adverse environmental temperatures negatively affect ectothermic animals' 51 temperature regulation time, reproductive strategy, gender ratio, and survival ability of offspring 52 (Liu et al., 2022), ultimately leading to the extinction or evolution of a species. In the face of 53 changing environmental temperatures, reptiles mainly use two thermoregulatory methods, namely behavioral and physiological thermoregulation (Bels, 2019). Behavioral thermoregulation is the 54 55 primary temperature regulation strategy adopted by most lizards and is an important strategy for buffering climate fluctuations. 56

In addition to climate change, environmental contamination also threatens reptile populations,
resulting in the global decline of reptiles (Todd et al., 2010). Reptiles can be exposed to

59 environmental pollutants through several routes, including ingestion of contaminated material, 60 dermal contact, maternal transfer to eggs and embryos, and incubating egg uptake from nest sites (Rich and Talent, 2009). Exposure to toxic substances may impair an organism's ability to cope with 61 warming, while warming can also alter the effects of toxic substances on reptiles, as the physical 62 63 and bioactive properties of compounds change with temperature. This affects changes in their 64 environmental fate, bioavailability and degradation processes, and the sensitivity of organisms to 65 toxic substances will also change with body temperature. Since the commercial product Roundup® with glyphosate (GLY) as the active ingredient was 66 67 produced in 1974 (Vijay K. Nandula, 2010), glyphosate-based herbicides (GBH) have become 68 popular due to their strong non-selective and high herbicidal activity. Especially with the large-scale 69 planting of genetically engineered glyphosate-tolerant crops around the world, glyphosate dominate 70 the global herbicide market (Benbrook, 2016). However, the application of large quantities of glyphosate results in the accumulation of this chemical in farmland and surrounding soils and waters. 71 72 It may further enter the air, surface water, and groundwater through evaporation, leaching, and 73 surface runoff, thereby threatening biodiversity and ecological safety. Residues of glyphosate and 74 its metabolite aminomethylphosphonic acid (AMPA) have been detected in soils in different regions 75 all over the world (Laitinen et al., 2009; Maggi et al., 2020; Pelosi et al., 2022; Wee et al., 2021). 76 Glyphosate and its commercial formulations have adverse effects on environmental organisms such 77 as soil-living animals, aquatic organisms, amphibians, insects, and birds through oxidative stress, 78 genotoxicity, neurotoxicity and growth toxicity (Gill et al., 2018). In addition, commercial 79 formulations are actually thought to be more toxic than the glyphosate active ingredient alone. The 80 physiological stress caused by glyphosate herbicides is a new challenge for reptiles, which have

82

emerging stress.

evolved over hundreds of millions of years, and it is not known how reptiles will respond to this

83 Chemical pollution and global warming are two major threats to reptiles, and these two factors can 84 interact with each other. This constitutes a key challenge for the realistic ecological risk assessment 85 and biodiversity research on the effects of global warming on reptile populations. Different studies 86 have demonstrated that reptiles can cope with global warming in various plastic ways, but exposure 87 to glyphosate herbicide pollution poses new challenges to the ability of lizards to resist climate 88 change-related stressors. We investigated thermoregulatory patterns in lizards in different thermal and glyphosate-exposed environments and explored the underlying mechanisms behind their altered 89 90 thermoregulation. We hypothesized that: (1) Since reptiles, as the first truly terrestrial vertebrates 91 to adapt to terrestrial life, have various plastic abilities to adapt to complex and changeable land 92 habitats, small increases in ambient temperature would not pose a significant threat to these lizards, 93 and that they would alter their physiological state and behavioral activities to respond effectively to 94 thermal challenges; (2) within the GBH exposed habitats, the toxic effects of GBH will result in 95 impaired fitness and loss of the ability to precisely regulate body temperature in response to 96 increasing temperatures.

- 97 **2. Material and methods**
- 98 2.1 Chemical reagents

99 The Glyphosate-based herbicide, -Roundup used in this study, was purchased from local stores 100 (Beijing. China). The analytical-grade glyphosate (99.5%) was provided by Shenyang Research 101 Institute chemical industry, AMPA (99%) obtained from Macklin of was 102 Biochemical Technology Co., Ltd (Shanghai). 4-chloro-7-nitro-1,2,3-benzoxadiazole (NBD-Cl,

103 98%) and other reagents were obtained from Sigma-Aldrich Co. (USA).

104 **2.2 Experimental animals and husbandry**

105 Eremias argus is a small live bearing lacertid lizard species inhabiting grasslands and has been designated an endangered species by the Korean Ministry of Environment since 2005(Kim et al., 106 107 2010). They are widely distributed, ranging from the Korean peninsula, Mongolia to certain areas 108 of Russia and China (Zhao et al., 1999). They commonly hibernate from October to March. Adult 109 female lizards were randomly selected from a semi-natural colony at China Agricultural University 110 after emergence and transferred to the laboratory. Lizards were housed in plastic incubators (58 \times 111 41×31 cm) with a layer of sand at the bottom. Water was supplied in glass containers situated randomly throughout the incubator, and food (mealworms and crickets sprinkled with vitamin 112 113 powder) was provided daily. Each terrarium was divided into two equal-sized areas: one area with 114 a reflector bulb (25 W, Reptile Basking Spot Lamps;) hanging directly above it, is hereafter referred 115 to as the "warm" sector, while the other area supplied paper egg trays as shelters would be referred 116 to as the "cool" sector (Figure S1). The heating bulbs were sequentially on for 8 h/day between 117 09.00 and 17.00. Besides, LED lights provided a 12h light: 12h dark cycle without heat. The lizards were allowed to adapt to the experimental environment for one month before the start of the 118 119 experiment.

120 2.3 Experimental design

121 The experimental design had two treatment factors: GBH exposure (with or without GBH exposure)
122 × two temperatures (current climate treatment and warmer climate treatment). We altered the
123 ambient temperatures of the incubators by using reflector bulbs with two different intensities for the
124 two thermal treatments: one with a 25 W bulb above the terrariums, termed the current climate

125	treatment (CT), and another treatment with a 50 W bulb above the terrariums, termed the warmer
126	climate treatment (WT). Electronic temperature data loggers (Apresys) were used to record the
127	temperature of the experimental terrarium every 20 minutes. Air temperatures were registered at 5
128	cm above the substrate in the middle of incubators. Soil temperatures mean the direct substrate
129	temperature in cool or warm sectors. Initial analyses showed that the mean air temperatures in CT
130	and WT terrariums were 26.99 and 30.18°C, respectively. We established a thermally heterogeneous
131	microhabitat within which the lizards can thermoregulate themselves in each treatment (Figure S2,
132	Table S1).
133	We contaminated the soil with 18 mg of glyphosate per kg soil. This method and concentration are
134	consistent with many ecotoxicology experiments on earthworms when these products are used as
135	directed. The commercial formulation of glyphosate (Roundup®) contains glyphosate (30%), The
136	recommended application of Roundup to field is 7.5L/ha, which equals 0.225ml a.i/m ² of glyphosate.
137	We used twice as much as the label suggested to kill weeds in an area the size of our incubator, and
138	added the predetermined 0.36 mL of Roundup® to the appropriate incubators with 6kg soil sand,
139	mixing the soil thoroughly. This resulted in a nominal initial concentration of approximately 18 mg
140	of glyphosate per kg soil (twice the recommended dose). This amount is within the normal range of
141	soil glyphosate concentrations used in most published toxicological studies.
142	Thus, there were 4 treatments in the study: Control group (CT-CK), GBH polluted soil group (CT-
143	GBH), high temperature without GBH soil group (WT-CK), and high temperature with GBH
144	polluted soil group (WT-GBH) (Table S2). All lizards were randomly separated into four groups
145	(18 females per group) for 60 days (lizards were separated into 12 plastic incubators, three

146 incubators for each group with three replicates, and six females per incubators).

After exposure for eight weeks, lizards were sacrificed by rapid decapitation after anesthesia. Their brain and liver tissue were dissected, weighed, and collected in 2 mL tubes. Blood was collected in an EDTA-2K tube, and plasma was collected by centrifugation. Samples for chemical or biochemical assays were stored at -80 °C until further use. Animal experiments and protocols were conducted under strict adherence to the principles for good laboratory animal care and were followed the ethical guidelines set by China Agricultural University.

153 **2.4 Behavioral observation**

The thermal microhabitat preference (i.e., which sector lizards were observed in) and behavioral 154 155 patterns of lizards were evaluated at the end of the experiment (either after 59 or 60 days). A digital 156 camera (HERO8, GoPro Inc., capturing one frame every two seconds) was used to photograph the 157 lizard from above during their active period (from 09:00 to 17:00). The image sequences were 158 extracted every 5 min to examine the behavior of lizards. The lizard's location and behaviors were classified into the following patterns: (1) cold sector: the lizard was hiding in the shelter (invisible) 159 160 or immobile on the ground; (2) moving: walking across the terrarium (noticeable ghosting and 161 blurry profiles of lizards can be observed because of fast-paced walking or running); (3) immobile 162 in the warm sector: An inactive lizard (not moving, scratching or basking) was defined as being 163 immobile; (4) basking in the warm sector: lizards were basking under the heat lamp (upright head 164 position and increased respiration). Finally, we counted the number of times each lizard exhibited a 165 given behavior.

166 **2.5 Thermal ecology parameters**

167 **2.5.1 Body temperature in the micro-environment**

168 To avoid disturbance caused by the repeated handling and cloacal body temperature measurements,

169 we measured the temperature of the body surface of the experimental lizards as body temperature

- 170 (T_b) using AR320 non-contact infrared thermometer (SMART SENSOR®) (on the 57th day). The
- 171 lizards were individually marked by paint codes before monitoring their active body temperature.
- 172 During the last days of experiment, we recorded the body temperature of all individuals from 10:00
- to 16:00 hours at hourly intervals. If a lizard was found to be inactive, we gently lifted the refuge
- 174 and measured lizards' body temperature. If they were hidden in the substrate and we could not locate
- them, body temperature measurements were left unscored.

176 **2.5.2 Preferred body temperature**

177 Ectotherms' preferred body temperature (T_{set}) is that one they attempt to achieve. It is calculated as

178 the mean of T_b selected by organisms free to choose where to thermoregulate in a laboratory gradient.

- 179 To determine T_{set}, we constructed a laboratory thermal gradient adapted from previous literature
- 180 (Wang et al., 2021). The device is shown in the Fig. S3, focused on providing an ecologically cost-
- 181 reduced environment aimed to lessen adverse effects of thermoregulation. On the 58th day, each
- 182 lizard was placed on a track for 2 h acclimation, and then the body temperature was measured every
- 183 10 min using AR320 non-contact infrared thermometer (SMART SENSOR®).
- 184 **2.5.3 The thermoregulatory accuracy**
- 185 We estimated the accuracy of thermoregulation (d_b) as the absolute value of the difference between
- 186 the T_{set} and the field-active T_b of animals ($d_b = |T_b T_{set}|$). Low values of d_b represent high
- 187 thermoregulatory accuracy because animals can achieve their T_{set} or close to it in the field. In
- 188 contrast, high values of d_b represent low thermoregulatory accuracy because T_b is either much higher
- 189 or lower than T_{set} .

190 **2.6 GBH analysis**

	191	Soil (20 g) was	collected at 1,	3, 7,	14, 28,	and 60 day	ys for soil o	degradation d	ynamics of C	JLY and
--	-----	-----------------	-----------------	-------	---------	------------	---------------	---------------	--------------	----------------

- 192 AMPA. The contents of GLY and AMPA were analyzed by high-performance liquid
- 193 chromatography (HPLC) using a modification of the method proposed by Wu et al. (2020) (Wu and
- 194 Lin, 2020). The recoveries of GLY and AMPA in the soil are listed in Table S3&S4.
- 195 **2.7 Biological analysis**
- 196 **2.7.1 Assay of enzyme activity**
- 197 Commercial test kits for Glutathione S-transferase (GST), superoxide dismutase (SOD), Malondialdehyde (MDA) were purchased from Elabscience Biotechnology Co., Ltd (Wuhan, 198 199 China). Assay kits for glucuronosyltransferase (UDPGT) and cytochromes P450 (CYP450) were 200 purchased from Jiangsu Suke Biotechnology Co., Ltd (Jiangsu, China). The customized enzyme-201 linked immunosorbent assay (ELISA) kits for reptilian heat shock protein 70 (HSP70) were 202 generated by Nanjing Jiancheng Bioengineering Institute (Nanjing, China). Tissue samples were homogenized in ice-cold phosphate buffer (0.01 M, pH 7.4) at 4 °C, and a 10% (w/v) homogenate 203 204 was centrifuged for 4000 rpm, 15 min (4 °C). The obtained supernatant was used to measure GST 205 and SOD activities, as well as MDA, UDPGT, CYP450 and HSP70 content.
- 206 2.7.2 Metabolomics

Lizard brain sample extraction was performed as previously described with some modifications(Chang et al., 2022; Zhang et al., 2018). Details are summarized in Supplementary Material. The original result file obtained by instrument analysis was imported into Compound Discoverer 3.2 software (Thermo Fisher Scientific, USA) for metabolites identified. Subsequently, the on-line software MetaboAnalyst 5.0 was used to screen differentially expressed metabolites (DEMs). DEMs was based on a VIP value > 1 for the first principal component of the OPLS-DA

213	model and a p-value < 0.05 for the T test. The SCMs were used for the pathway analysis via
214	MetaboAnalyst 5.0. The pathways with p value < 0.05 were considered as potentially affected.
215	2.7.3 RT-qPCR
216	Roughly 50mg liver samples and 50mg brain samples were homogenized in TRNzol Universal
217	reagent (TIANGEN Biotech Co., Ltd) for total RNA isolation, following the manufacturer's
218	instructions. Total RNA concentration was calculated from the absorbance at 260 nm using a
219	NanoDrop-1000 spectrophotometer (NanoDrop Technologies Inc.). To ensure RNA purity, quality
220	was verified by gel electrophoresis in 1% agarose gel, and only RNA samples with OD260/OD280
221	between 1.7 and 2.0 were used for cDNA synthesis. Complimentary DNA (cDNA) was generated
222	from 1 µg total RNA using FastKing RT Kit (with gDNase), as described by the manufacturer
223	(TIANGEN Biotech Co., Ltd). The quantitative real-time polymerase chain reaction (RT-qPCR)
224	were carried out using the TIANGEN RT-qPCR kits (Biotech, Beijing, China). The genes of
225	members of the thermosensitive transient receptor potential ion channels family (Thermo-TRPs)
226	were selected for gene expression analysis and their primers are listed in Table S5. The RT-qPCR
227	was carried out with Bio-Rad CFX (Bio-Rad, USA). The thermal cycling program was operated
228	according to the following protocol: 95 °C for 15 min, followed by 40 cycles of 95 °C for 10 s,
229	60 °C for 20 s and 72 °C for 20 s. The β -actin housekeeping gene was used for the internal
230	normalization of the target genes. The relative expression was calculated using b-actin and the
231	results were analyzed according to the delta-delta Ct method ($\Delta\Delta$ Ct).

232 2.8 Data analyses

The results were expressed as mean ± standard deviation (mean ± SD) and analyzed with SPSS
v20.0 (IBM, USA). Graphical plots were realized with GraphPad Prism 9.5.0 (GraphPad Software,

235	Inc. USA). Data were tested for GBH and warming treatment intera	action effects by two-way
236	analysis of variance (ANOVA) followed by the post-hoc Duncan t	est. Statistical differences
237	between different treatments were conducted by One-way ANOVA wi	th a post-hoc Duncan test.
238	The assumptions of normality and homogeneity of variance were de	etermined for analysis via
239	Kolmogorov-Smirnov and Levenes test.	
240	3.Results	

241 3.1 Mortality

242 No mortality was observed in the control or exposure groups during the entire experimental period.







Fig. 1 Thermal performance parameters data of *Eremias argus* lizards. A: Partitioning of time among the different behaviors in lizards; B: The active body temperature of lizards; C: The preferred body temperature of lizards; D: The thermoregulatory accuracy of lizards; bars indicate standard deviation (SD). Boxplots provide the median (black horizontal line), the interquartile range

- 249 (upper and lower sides of the box) and minimum and maximum values (whiskers). Letters indicate
- 250 statistically significant differences between treatment groups.
- 251

252 Table 1 Two-way ANOVA analysis results of lizard thermal parameters and enzyme activity of

253	E	i.argus.				
	GE	BH	Temp	erature	GBH×Temperature	
variables	F	Р	F	Р	F	Р
The active body temperature (T _b)	6.488	0.014	27.710	<0.001	6.599	0.013
The preferred body temperature (T_{set})	13.081	0.001	1.430	0.237	0.048	0.828
The thermoregulatory accuracy (d_b)	7.775	0.008	0.026	0.873	0.781	0.381
GST	9.709	0.005	12.841	0.002	0.175	0.680
UDPGT	13.664	0.001	19.107	<0.001	0.010	0.919
CYP450	16.924	0.001	12.571	0.002	0.282	0.601
MDA	3.076	0.095	6.165	0.022	3.745	0.067
SOD	27.550	<0.001	65.615	<0.001	35.948	<0.001
HSP70 in liver	36.180	<0.001	5.909	0.025	1.020	0.325
HSP70 in plasma	1.719	0.205	5.585	0.028	1.181	0.290

Through the observation of the daily behavioral patterns of lizards, it was found that compared with other groups of lizards, the proportion of basking in WT-CK group was significantly reduced (Fig. 1A). We found that both GBH and ambient temperature each had a significant effect on T_b in the micro-environment ($F_{1,54}$ =6.488, P=0.014 ; $F_{1,54}$ =27.710, P<0.001), also with a significant interaction effect ($F_{1,54}$ =6.599, P=0.013) (Table 1). According to post-hoc tests, simultaneous exposure to GBH and higher ambient temperature resulted in a significant increase in T_b in the lizards, while there was no significant difference in the T_b of the other three groups of lizards (Fig.

262 1B). We also found that GBH had a significant effect on T_{set} (F_{1,58}=13.081, P=0.001), while 263 environmental temperature had no significant effect on T_{set} (F_{1,58}=1.430, P>0.05) and there was no interaction effect between the two variables (F_{1.58}=0.048, P>0.05) (Table 1). In the current climate 264 265 treatment groups, lizards which exposed to GBH preferred higher body temperatures relative to the 266 control group. GBH exposure also resulted in a significant increase in preferred body temperature 267 in the warmer climate treatment groups (Fig. 1C). As for the thermoregulatory accuracy of the lizards: GBH had a significant effect on thermoregulatory accuracy (F_{1.47}=7.775, P=0.008), ambient 268 temperature had no significant effect on thermoregulatory accuracy (F1,47=0.026, P>0.05) and there 269 was no significant interaction between the two variables ($F_{1,47}=0.781$, P>0.05) (Table 1). In the 270 271 current climate treatment, GBH exposure led to significantly higher in d_b values, suggesting that 272 GBH decreased in the lizard's ability to precisely regulate its body temperature. However, the same 273 results were not observed in the two warmer climate treatment groups (Fig. 1D).

274 **3.3 Degradation of glyphosate in soils**



275



The soil degradation dynamics of GLY and AMPA over 60 days were determined and results are shown in Fig. 2 and Table 2. The nominal exposure concentrations of GLY in the soil were 19.49 mg/kg and 18.95 mg/kg in the CT-GBH and WT-GBH groups, respectively. There was some

degradation of glyphosate in the soil during the 60-day exposure period, with a degradation rate of

282 29.8% in the CT-GBH groups, and more rapid degradation of 37.2% in the WT-GBH group. Its

283 metabolite AMPA was also detected in the soil at the concentrations shown in Fig. 2.

284	Table 2. The	concentration	and deg	adation ra	te of GLY	and AMPA	in the soil	of incubat	tors
-----	--------------	---------------	---------	------------	-----------	----------	-------------	------------	------

Treatment	GLY concentration in soil		Degradation	AMPA concentration in soil	
group	(mg/kg soil)		rate (%)	(mg/kg soil)	
	0 d	60 d		0 d	60 d
CT-CK	ND	ND	-	ND	ND
CT-GBH	19.49±1.95	13.69±2.10	29.8	0.65±0.031	2.10±0.05
WT-CK	ND	ND	-	ND	ND
WT-GBH	18.95±1.68	11.90±2.39	37.2	1.20±0.11	3.13±0.23

285

286 **3.4 Response of detoxification enzyme activities**



287

288 Fig. 3 Effects of GBH exposure and differential temperature treatments on the liver GST, UDPGT,

and CYP450 of lizards. Letters indicate statistically significant differences between treatmentgroups.

291 The liver GST activity, UDPGT, and CYP450 contents of the lizards under each treatment are 292 shown in Fig. 3, and the results of the two-way ANOVA analysis are shown in Table 1. The three

293	biochemical parameters mentioned above did not interact significantly with GBH or ambient
294	temperature. Both temperature and GBH lead to a significant increase in these three enzymes.
295	Without glyphosate exposure, liver GST activity, UDPGT, and CYP450 content increased with the
296	increase of ambient temperature (CT-CK vs WT-CK). At the current climate treatment, GBH
297	exposure resulted in an increase in liver GST activity, UDPGT, and CYP450 by 26.62%, 18.14%,
298	and 21.31%, respectively (CT-GBH vs CT-GBH). In the warmer climate treatment groups, liver
299	GST activity, UDPGT, and CYP450 contents were 26.57%, 15.78%, and 23.42% higher in the
300	GBH-treated group than that of the control group, respectively (WT-GBH vs WT-GBH).

301 **3.4 Thermal stress and antioxidant system**





Fig. 4 Effects of GBH exposure and differential temperature treatments on brain MDA, SOD, and
 HSP70 levels in the liver and plasma of lizards. Letters indicate statistically significant differences

305 between treatment groups.

306

307	Exposure to GBH significantly increased the MDA content in lizard brain tissue, independent of
308	changes in ambient temperature (GBH \times Temperature: F _{1,20} =3.745, P=0.067; GBH: F _{1,20} =3.076,
309	P=0.095; Temperature: $F_{1,20}$ =6.165, P=0.022). The level of MDA in the brain tissue of lizards in the
310	CT-GBH group was significantly higher than that of the other three groups, and there was no
311	significant difference between the other three groups. Both GBH and ambient temperature had
312	significant effects on brain SOD enzyme activity (GBH \times Temperature: F _{1,20} =35.948, P<0.001;
313	GBH: $F_{1,20}$ =27.550, P<0.001; Temperature: $F_{1,20}$ =65.615, P<0.001). We also found that SOD
314	enzyme activity was significantly higher in the lizards in the WT-GBH group compared to the other
315	three groups, with no significant difference between the other three groups.
316	The expression of heat shock protein 70 (Hsp70) was used as a general stress response indicator to
317	monitor environmental stressors. According to the main effect model, temperature generally causes
318	liver (GBH × Temperature: $F_{1,20}$ =1.020, P=0.325; GBH: $F_{1,20}$ =36.180, P<0.001; Temperature:
319	$F_{1,20}$ =5.909, P=0.025) and plasma Hsp70 (GBH × Temperature: $F_{1,20}$ =1.181, P=0.290; GBH:
320	$F_{1,20}$ =1.719, P=0.205; Temperature: $F_{1,20}$ =5.585, P=0.028), and Hsp70 levels in lizard liver were
321	elevated by GBH exposure regardless of temperature.

322

3.5 Metabolomics study in brain

Non-targeted metabolomics analysis revealed that a total of 82 common metabolites were identified in the upper extracts of the four sets of brain tissue samples, and a total of 131 common metabolites were identified in the lower phase. The compounds extracted from the upper phase were mainly non-polar metabolites, like lipids and lipid-like molecules, while the lower phases were polar

327	metabolites, mainly nucleosides, nucleotides, and analogues, organic acids and their derivatives,
328	etc. Based on OPLS-DA model analysis and t-test analysis, there were 14 (upper) and 26 (lower)
329	differential metabolites in the lizard brains of the CT-GBH group compared to the CT-CK group
330	(Fig. 5A, D). The differential metabolites that were significantly altered in this group were used for
331	clustering analysis to generate heat maps, and these differential metabolites were used to examine
332	the changes in metabolic pathways via KEGG pathway analysis to further investigate the disruption
333	of metabolic pathways caused by glyphosate exposure. As shown in the Fig. 5J, exposure to GBH
334	can affect these metabolic pathways in the brain: histidine metabolism, beta-Alanine metabolism,
335	Aminoacyl-tRNA biosynthesis, Galactose metabolism, Amino sugar and nucleotide sugar
336	metabolism, Phenylalanine, tyrosine and tryptophan biosynthesis, Tyrosine metabolism. We
337	identified 15 (upper) and 16 (lower) metabolites in WT-CK group that changed significantly
338	compared with CT-CK group following an increase in ambient temperature (Fig. 5B, E). Compared
339	to the control, the warmer climate treatment had significant effects on some metabolic pathways:
340	Glutathione metabolism, Aminoacyl-tRNA biosynthesis, Sphingolipid metabolism, Alanine,
341	aspartate and glutamate metabolism, Pantothenate and CoA biosynthesis. Compared with CT-CK
342	group, the WT-GBH group had 14 (upper) and 18 (lower) differential metabolites in lizard brain
343	(Fig. 5C, F). Some pathways were significantly impacted between the WT-GBH and CT-CK groups,
344	including purine metabolism, Alanine, aspartate and glutamate metabolism, Pantothenate and CoA
345	biosynsis, Valine, leucine and isoleucine degradation, Arginine biosynsis, Aminoacyl-tRNA
346	biosynsis.



Fig. 5 The heat maps, categorical proportions and KEGG pathway analysis of differential metabolites identified from lizard brains. (A, D, G and J) CT-GBH vs CT-CK; (B, E, H and K) WT-

350 CK vs CT-CK; (C, F, I and L) WT-GBH vs CT-CK.

351 **3.6** Gene expression of thermosensitive transient receptor potential ion channels (thermo-

352 **TRPs**)

The gene expressions of the members of the thermosensitive transient receptor potential ion channels family in lizard brain and liver are shown in Fig.6. We found that compared with the CT-CK group, the expression of *TRPM7* in the liver of the lizards in the CT-GBH group was significantly decreased (p<0.05), while the expression of *TRPA1* was significantly increased, and the expression of *TRPM7*, *TRPA1*, and *TRPV2* in the brain tissue was down-regulated (p<0.05).

Compared with the control group, the warmer climate treatment significantly increased *TRPM7*, *TRPA1*, and *TRPV2* mRNA expression in the liver of the lizards in the WT-CK group, as well as the expression of *TRPM7* and *TRPA1* in the brain tissue. In the WT-GBH group, the expression of *TRPV2* was up-regulated in the liver and down-regulated in the brain.



Fig. 6 Effects of temperature and GBH exposure on the expression of genes related to thermosensitive transient receptor potential ion channels (thermo-TRPs) in the liver and brain tissues of lizards.

367 **4. Discussion**

368 4.1 The plastic response of lizards to warming

If the ambient temperature is higher than the ectotherm's optimal physiological range, several aspects of the behavior and physiology may be negatively affected, including locomotion, sensory input, foraging ability, rates of feeding and growth and immune function (Angilletta et al., 2002). Regardless of ambient temperature, endotherms can are gifted with an ability to regulate a steady temperature inside their bodies through metabolic heat production and loss processes (Boyles et al., 2011). Although ectotherms do not maintain body temperature homeostasis in a similar manner, they can regulate body temperature through specific thermoregulatory strategies, including behavior

and physiological thermoregulation (Kearney et al., 2009; Seebacher et al., 2015).

377 In this study, when there was no GBH exposure treatment, the active body temperatures measured 378 in lizards did not increase, even though the ambient temperature increased, and their accuracy of thermoregulation was not affected by warmer climate treatment. This suggests that *E.argues* lizards 379 380 can response to small increases in ambient temperature, and regulate their body temperature close 381 to their preferred body temperature and meets their heat needs. Adjustment of basking position and 382 posture, microhabitat selection and shuttling, and changes in basking frequency are considered to 383 be some important aspects of behavioral thermoregulation in these reptile species (Black et al., 2019; 384 Sinervo et al., 2010b; Spellerberg, 1972). Similar to these, it was observed in this study that the 385 frequency of basking behavior decreased significantly after the temperature increased, while time spent on cold sector increased (although not statistically significant), which confirmed that they 386 387 compensated for the change of thermal environment by altering basking frequency and microhabitat 388 use. Furthermore, T_{set} of lizards was measured in our laboratory setup with a temperature gradient 389 without ecological constraints. The wide range of T_{set} in the CT-CK and WT-CK groups exemplifies 390 that lizards, as habitat generalists and ambush foragers, experience a complex and variable 391 environment, and exhibit more eurythermic behaviors and are adapted to a variety of environmental 392 temperatures, with a buffering capacity to cope with some degree of fluctuating environmental 393 temperatures (Crowell et al., 2021). 394 Heat shock proteins (HSPs) are a super-family of proteins in response to various biological and

abiotic stressors, also known as stress proteins (de Pomerai, 1996; Feder and Hofmann, 1999; Hamer
et al., 2004). The physiological response to hyperthermia exposure is often accompanied by the
expression of HSPs, which participates in protecting cells from heat damage by protecting cell
proteins from denaturation (Hassan et al., 2019). The level of Hsp70 in the liver and plasma of

399	lizards in WT-CK group increased significantly with ambient temperature, which is considered a
400	protective mechanism against heat stress. For example, <i>rainbow trout</i> has a time- and tissue-specific
401	Hsp30 response to heat stress, which is expected to provide some protection in heat stress within a
402	certain range (Liu et al., 2019). Also, HSP70 can promote thermal tolerance of reptiles embryos and
403	plays an important role in coping with acute heat stress in koi (Cyprinus carpio) (Gao et al., 2014;
404	Yu et al., 2020). HSPs are ecologically relevant mechanisms used by animals to promote thermal
405	adaptation and provide protection to organisms in heat stress (Jin et al., 2020; Kaur et al., 2011).
406	It has been shown that the brain is the most sensitive organ to thermal stimuli in Cuban Anolis
407	lizards(Akashi et al., 2016). Several studies have suggested that physiological changes in the brain
408	may be related to Acclimatisation or acclimation responses to environmental temperatures. Lizards
409	(Calotes versicolor) adapt thermally by increasing the relative proportion of high melting point
410	phospholipids to alter the phase-transition temperature of hypothalamic cell membrane
411	phospholipids. In addition, the proportion of unsaturated fatty acids in the hypothalamus and the
412	brain as a whole increased with cold adaptation (Durairaj and Vijayakumar, 1984). These changes
413	in the fatty acid composition of the brain and hypothalamus would provide signals for temperature
414	selection or compensation (Seebacher, 2005). In this study, metabolomic analysis of lizard brain
415	tissue revealed that the majority of lipid metabolites up-regulated in brain tissue of the WT-CK
416	group following increased ambient temperature were of the phospholipid and sphingolipid classes
417	(Cer(d18:0/18:0), Phytosphingosine, Cer(d18:1/16:0), PE(P-16:0/20:3(8Z,11Z,14Z)), Sphingosine,
418	C14 Sphingomyelin, 1-Palmitoylphosphatidylcholine, LysoPE(18:0/0:0), Stearoyl
419	lysophosphatidylethanolamine). Phospholipids are one of the main components of cell membranes,
420	include phosphatidylcholine (PC), phosphatidylethanolamine (PE), phosphatidylserine (PS) etc.

421	Among them, PE and PC are the main components of neuronal cell membranes, and they are
422	involved in cellular signal transduction processes as precursors of second messengers (Gonzalez-
423	Riano et al., 2016; Joensuu et al., 2020). Studies have also proposed to adjust the fatty acid
424	composition of biological membranes to achieve temperature adaptation in ectotherms (Price et al.,
425	2017). The phospholipid content in lizards' brain increases in warmer environments in this study,
426	to allow temperature acclimation by maintenance of neuronal membrane stability and fluidity.
427	Sphingolipids are lipids that are closely associated with the nervous system and they play an
428	important role in maintaining the normal structure and function of the nervous system. In our study,
429	the changes in brain phospholipid and sphingolipid content in the WT-CK group might assist the
430	lizard in coping with the increased ambient temperature.
431	In ectotherm, thermo sensory organs receive thermal information, then send it to and integrate with
432	the central nervous system (CNS) which modulate behavioral outputs in a top-down fashion (Abram
433	et al., 2017). Various membrane ion channels and receptors act as the initial detectors of
434	environmental temperature cues. Among them, thermosensitive transient receptor potential ion
435	channels (thermo-TRPs) are arguably the most well-established molecular thermosensors in cross-
436	species studies (Xiao and Xu, 2021). For example, TRPV1 and the cool-sensing TRPM8 are
437	expressed in tissues throughout the body of the crocodile (Crocodylus porosus). Inhibition of TRP
438	proteins in C. porosus can significantly alter thermoregulatory behaviour, abolishes shuttling
439	behaviour, and leads to significantly altered body temperature patterns (Seebacher and Murray,
440	2007). In Cuban Anolis lizards, thermal tolerances are related to transient receptor potential ion
441	channel ankyrin 1 (TRPA1) heat sensitivities (Akashi et al., 2018). As TRPs are present in reptiles
442	and function to control thermoregulatory behaviour. We detected the change of mRNA expression

of *TRPV2*, *TRPM7* and *TRPA1* which are the subfamilies of this family. We found that all three showed increased transcription and function in lizard brain tissue in response to higher ambient temperatures. Taken together, the lizard can adjust its physiological status (Hsp70 content, brain lipid changes, mRNA expression) along with behavioral strategies, to adapt to the rise of ambient temperature.

448 **4.2 Interaction between temperature and glyphosate**

449 A previous study found that the fatal toxicity of Beta-cyfluthrin (BC) to lizards was negatively 450 associated with temperature. Therefore, after exposing to low BC dose, most lizards sought a 451 warming environment. The toxicokinetic process of toxicants (absorption, distribution, metabolism 452 and clearance) in animals is temperature-dependent (Brown et al., 2021). Among them, the 453 metabolism of exogenous pollutants is a process of enzyme-catalyzed biotransformation where the 454 rate of enzyme-catalyzed reaction is determined by the amount of enzyme and the temperature. It is also known that relatively high temperatures are conducive to the biotransformation of toxicants. 455 456 Therefore, the thermoregulation behavior of seeking warmer environment after BC exposure is 457 considered as a "self- rescue" behavior in lizards (Wang et al., 2022). In this study, whether treated with the current climate or warmer climate conditions, GBH lead to an increase in T_{set} of lizard (CT-458 459 GBH, WT-GBH), meaning they actively seek warming. Such thermoregulation behavior is a 460 survival strategy in response to GBH exposure stress and increases the possibility of their survival. 461 Similarly, the dermal exposure of the diurnal lizard (Oligosoma polychroma) to Yates Roundup 462 Weedkiller (commercial glyphosate formulations) caused no change in body mass and a significant increase in selected temperatures (Carpenter et al., 2016). The potential mechanism underlying this 463 heat-seeking behavior may be to increase the fever response of metabolism, thus offsetting 464

465 physiological stress.

Body temperature can aid reptiles to better cope with environmental pressure by increasing the 466 effectivity of physiological and biochemical systems because high body temperatures can activate 467 biological conversion enzymes which affect the metabolism and elimination of the poison. The T_b 468 469 of the WT-GBH lizard was significantly higher than other groups, indicating that they regulate the 470 body temperature so that it is closer to their prefered body temperature while improving the metabolic rate to facilitate detoxification at the same time. On the contrary T_b in CT-GBH lizards 471 472 has not improved. The liver through biological metabolism and detoxifying exogenous compounds (Kedderis, 2010). This process is known as Phase I and Phase II metabolism. The enzymes 473 participating in this process include the cytochrome P450 enzymes, UDG-PT and GST (Rourke and 474 475 Sinal, 2014). In our study, GBH and warmer temperature treatment could activate CYP450, UDGPT 476 and GST, and levels of the three enzymes in the WT-GBH lizards are significantly higher than in the other three groups. High body temperature can protect liver tissue by stimulating the production 477 478 of liver metabolic enzymes, thus eliminating endogenous or exogenous harmful substances faster 479 and more efficiently. 480 In addition, temperature can not only change the physiological characteristics of animals, but also 481 change the behaviour and environmental fate of pesticides (Bento et al., 2016; Broznić and Milin, 482 2012; Daam and Van den Brink, 2010). Our results show that glyphosate degrades faster in soil 483 under warmer environment. Glyphosate is mainly decomposed by microorganisms, and microbial 484 activity may increase at higher temperatures (Sviridov et al., 2015; Zabalov et al., 2022). An 485 earthworm's (Eisenia fetida) sensitivity to glyphosate varies with specific environmental conditions, 486 where glyphosate toxicity decreased with an increase in soil temperature (Pochron et al., 2019).

Developmental acclimation to higher temperatures reduces the sensitivity of Ceriodaphnia silvetrii 487 488 to glyphosate (Silva et al., 2020). In an aquatic environment, a 0.5-degree temperature increase renders glyphosate toxicity more severe in zebrafish(Sulukan et al., 2023). During glyphosate 489 490 degradation, the metabolite aminomethyl phosphonic acid (AMPA) will be produced. Few studies 491 have shown that toxicological effect of AMPA to reptiles. The toxicity of glyphosate also depends 492 on other components in the commercial preparation of Roundup[®] (Gill et al., 2018), which was not 493 included in this study. We suggest that the toxicity effects of AMPA and other components in the 494 formulations on lizards should be taken into account further.

Higher ambient temperature can lower the overall risk of GBH as high temperature has been associated with the greater dissipation of GLY and the increase of biochemical detoxification mechanism by lizards. All in all, the lizards in WT-GBH group were not subject to GBH toxicity stress, and the ability to control body temperature is not affected by GBH. The degradation of glyphosate in the environment is accelerated with the increase of temperature, the activity of biotransformation enzyme is high, and the metabolism and elimination of toxicants are fast.

501 **4.3 The thermoregulation of lizards affected by GBH**

Metabolic functions are usually optimized within a narrow range of body temperatures (Tb's), making thermoregulation essential for the survival and fitness of an animal. However, thermoregulation behavior of reptiles has associated costs. The cost-benefit model predicts that only when the related fitness costs (such as time or energy loss, and predation risk) do not exceed the fitness benefits, can the ectotherms take the initiative to carry out beneficial thermoregulatory adjustments (Huey and Slatkin, 1976). This strategy enhances fitness by allowing lizards to adapt more flexibly to changing environmental conditions, while maximizing the benefits of maintaining 509 high Tb and minimizing energy consumption. A study evaluated the thermoregulatory behavior of inland bearded dragons (Pogona vitticeps) under various thermal quality conditions requiring 510 different locomotory investments, and it showed that thermoregulatory precision was decreased in 511 512 environments where the motor costs associated with thermoregulation were high (Cadena and 513 Tattersall, 2009). Another experimental study on the desert iguana (Diposaurus dorsalis) also 514 concluded that lizards that have to shuttle more frequently to regulate their body temperature will allow their body temperature passively close to the ambient temperature (Withers and Campbell, 515 516 1985). When more metabolic costs are required for body temperature regulation (energy cost of 517 shuttle movement), lizards were less able to regulate their body temperature. In this study, when 518 lizards were exposed to GBH at the current climate treatment, the variance of the T_b in lizards was 519 great, indicating that the thermoregulation is negative and not precise (Brattstrom, 1965); They also 520 did not frequently shuttle through microhabitats and increase the frequency of basking, possibly due to the high cost of accurate thermoregulation. Also, the lizards in the CT-GBH group may tend to 521 522 invest more energy in the detoxification metabolism of GBH, as it is a process that consumes a lot 523 of energy. When animals have a certain amount of energy, priority should be given to meeting the 524 life history characteristics of self-sustainment that require low level energy (such as oxidative stress, 525 detoxification metabolism, immune response) rather than higher-level energy-demanding life history activities such as fat storage, locomotor performance, reproduction, etc. (Rutschmann et al., 526 2021). 527 In vertebrates, neutral centers in the hypothalamus are responsible for the control of body 528

530 conducted a non-targeted metabolomic analysis of brain tissue and found abnormal histidine

529

temperature. To investigate the effect of GBH on the thermoregulatory of these lizards, we

531	metabolism in the CT-GBH lizards. Histidine can be enzymatically decarboxylated to obtain
532	histamine, which acts as a neurotransmitter in the brain and is associated with a variety of brain
533	functions, such as sleep-wakefulness, circadian rhythm, appetite, memory and stress responses
534	(Brosnan and Brosnan, 2020; Wada et al., 1991). The regulation of body temperature is largely
535	dependent on the histaminergic and cholinergic neuronal systems in the brain (A et al., 2010). The
536	existence of histaminergic thermoregulatory mechanisms in mice has been verified (Itoh et al.,
537	1986). The specific mechanism of histidine metabolism in reptile thermoregulation still needs
538	further research to clarify. Even though the lizards of the CT-GBH group preferred higher body
539	temperatures in the laboratory gradient, they did not refine their thermoregulation to bring their T _b
540	closer to the more beneficial high temperatures. On the one hand, the energy cost of the shuttle and
541	basking period was high. On the other hand, GBH caused oxidative damage to the brain tissue and
542	a significant disturbance in histidine metabolism of the CT-GBH lizards, possibly resulting in a
543	decline in the accuracy of body temperature regulation. However, thermoregulation is a complex
544	process, and the mechanisms by which GBH affects this process deserve further investigation. Over
545	a 60-day exposure to GBH, no mortality has been observed in any group. However, oxidative stress
546	levels in GBH-treated lizards indicate the presence of physiological toxic effects caused by GBH,
547	and as exposure time increases and climate change progresses, subtle toxicological effects may
548	increasingly limit thermoregulatory behavior and have a greater impact on the population as a whole.
549	5. Conclusion
550	With the development of global climate change, how the interaction of multiple stressors affects
551	reptiles is an urgent research focus. In this study, the WT-CK lizards thermoregulated with high
552	accuracy, which suggested that this species adapted to moderate environmental temperature increase

553	through thermoregulatory strategies. However, GBH exposure resulted in the lizard's
554	thermoregulatory accuracy decreased, but the lizards were protected from the adverse effects of
555	GBH through several temperature-dependent detoxification mechanisms facilitated by ambient
556	temperature rise. The subtle toxicological effects of GBH may increasingly limit the behavioral
557	response of this species more and more. Therefore, it is worth exploring further whether lizards may
558	be challenged to survive as the extreme climate situation increases sharply in the global warming
559	scenario, as well as the relationship between environmental pollution and animal thermal physiology.
560	

561 Acknowledgements

562 This work was supported by fund from the National Natural Science Foundation of China (Contract

563 Grant number: 42277276). Approval All experiments and protocols followed ethical guidelines set

564 by China Agricultural University.

565

567 References

- 568 A, T., G, N., Ib, M., O, E., 2010. Effects of anti-histaminic and anti-cholinergic substances on human
- thermoregulation during cold provocation. Brain Research Bulletin 81, 100-106.
- 570 https://doi.org/10.1016/j.brainresbull.2009.06.012
- 571 Abram, P.K., Boivin, G., Moiroux, J., Brodeur, J., 2017. Behavioural effects of temperature on
- 572 ectothermic animals: unifying thermal physiology and behavioural plasticity. Biological
 573 Reviews 92, 1859–1876. https://doi.org/10.1111/brv.12312
- 574 Akashi, H.D., Cádiz Díaz, A., Shigenobu, S., Makino, T., Kawata, M., 2016. Differentially
- expressed genes associated with adaptation to different thermal environments in three
 sympatric Cuban *Anolis* lizards. Mol Ecol 25, 2273–2285.
 https://doi.org/10.1111/mec.13625
- 578 Akashi, H.D., Saito, S., Cádiz Díaz, A., Makino, T., Tominaga, M., Kawata, M., 2018. Comparisons
- 579 of behavioural and TRPA1 heat sensitivities in three sympatric Cuban *Anolis* lizards. Mol

580 Ecol 27, 2234–2242. https://doi.org/10.1111/mec.14572

Angilletta, M.J., Niewiarowski, P.H., Navas, C.A., 2002. The evolution of thermal physiology in
 ectotherms. Journal of Thermal Biology 27, 249–268. https://doi.org/10.1016/S0306-

583 4565(01)00094-8

- Bels, V.L. (Ed.), 2019. Behavior of lizards: evolutionary and mechanistic perspectives. CRC Press,
 Boca Raton, Florida.
- Benbrook, C.M., 2016. Trends in glyphosate herbicide use in the United States and globally.
 Environmental Sciences Europe 28, 3. https://doi.org/10.1186/s12302-016-0070-0
 Bento, C.P.M., Yang, X., Gort, G., Xue, S., van Dam, R., Zomer, P., Mol, H.G.J., Ritsema, C.J.,

589	Geissen, V., 2016. Persistence of glyphosate and aminomethylphosphonic acid in loess soil
590	under different combinations of temperature, soil moisture and light/darkness. Science of
591	The Total Environment 572, 301–311. https://doi.org/10.1016/j.scitotenv.2016.07.215
592	Black, I.R.G., Berman, J.M., Cadena, V., Tattersall, G.J., 2019. Behavioral Thermoregulation in
593	Lizards: Strategies for Achieving Preferred Temperature, in: Behavior of Lizards. CRC
594	Press.
595	Boyles, J.G., Seebacher, F., Smit, B., McKechnie, A.E., 2011. Adaptive Thermoregulation in
596	Endotherms May Alter Responses to Climate Change. Integrative and Comparative
597	Biology 51, 676-690. https://doi.org/10.1093/icb/icr053
598	Brattstrom, B.H., 1965. Body Temperatures of Reptiles. The American Midland Naturalist 73, 376-
599	422. https://doi.org/10.2307/2423461
600	Brosnan, M.E., Brosnan, J.T., 2020. Histidine Metabolism and Function. The Journal of Nutrition
601	150, 2570S-2575S. https://doi.org/10.1093/jn/nxaa079
602	Brown, C.T., Yahn, J.M., Karasov, W.H., 2021. Warmer temperature increases toxicokinetic
603	elimination of PCBs and PBDEs in Northern leopard frog larvae (Lithobates pipiens).
604	Aquatic Toxicology 234, 105806. https://doi.org/10.1016/j.aquatox.2021.105806
605	Broznić, D., Milin, Č., 2012. Effects of temperature on sorption-desorption processes of
606	imidacloprid in soils of Croatian coastal regions. Journal of Environmental Science and
607	Health, Part B 47, 779–794. https://doi.org/10.1080/03601234.2012.676413
608	Cadena, V., Tattersall, G.J., 2009. The Effect of Thermal Quality on the Thermoregulatory Behavior
609	of the Bearded Dragon Pogona vitticeps: Influences of Methodological Assessment.
610	Physiological and Biochemical Zoology 82, 203–217. https://doi.org/10.1086/597483

- 611 Carpenter, J.K., Monks, J.M., Nelson, N., 2016. The effect of two glyphosate formulations on a
- 612 small, diurnal lizard (Oligosoma polychroma). Ecotoxicology 25, 548-554.
- 613 https://doi.org/10.1007/s10646-016-1613-2
- 614 Chang, J., Pan, Y., Liu, W., Xie, Y., Hao, W., Xu, P., Wang, Y., 2022. Acute temperature adaptation
- 615 mechanisms in the native reptile species Eremias argus. Science of The Total Environment

616 818, 151773. https://doi.org/10.1016/j.scitotenv.2021.151773

- 617 Crowell, H.L., King, K.C., Whelan, J.M., Harmel, M.V., Garcia, G., Gonzales, S.G., Maier, P.H.,
- 618 Neldner, H., Nhu, T., Nolan, J.T., Taylor, E.N., 2021. Thermal ecology and baseline
- 619 energetic requirements of a large-bodied ectotherm suggest resilience to climate change.
- 620 Ecol. Evol. 11, 8170–8182. https://doi.org/10.1002/ece3.7649
- 621 Daam, M.A., Van den Brink, P.J., 2010. Implications of differences between temperate and tropical
- freshwater ecosystems for the ecological risk assessment of pesticides. Ecotoxicology 19,
- 623 24–37. https://doi.org/10.1007/s10646-009-0402-6
- de Pomerai, D., 1996. Review : Heat-shock proteins as biomarkers of pollution. Hum Exp Toxicol
- 625 15, 279–285. https://doi.org/10.1177/096032719601500401
- Durairaj, G., Vijayakumar, I., 1984. Temperature acclimation and phospholipid phase transition in
 hypothalamic membrane phospholipids of garden lizard, Calotes versicolor. Biochimica et
 Biophysica Acta (BBA) Biomembranes 770, 7–14. https://doi.org/10.1016/00052736(84)90066-X
- Feder, M.E., Hofmann, G.E., 1999. HEAT-SHOCK PROTEINS, MOLECULAR CHAPERONES,
 AND THE STRESS RESPONSE: Evolutionary and Ecological Physiology. Annual
 Review of Physiology 61, 243–282. https://doi.org/10.1146/annurev.physiol.61.1.243

0.55 0.50 , 0.50	633	Gao, J., Zhang,	W., Dang,	W., Mou, Y	Y., Gao,	Y., Sun,	BJ., Du	, WG.	, 2014.	Heat shock	prote
---	-----	-----------------	-----------	------------	----------	----------	---------	-------	---------	------------	-------

- 634 expression enhances heat tolerance of reptile embryos. Proc. R. Soc. B. 281, 20141135.
- 635 https://doi.org/10.1098/rspb.2014.1135
- 636 Gill, J.P.K., Sethi, N., Mohan, A., Datta, S., Girdhar, M., 2018. Glyphosate toxicity for animals.
- 637 Environ Chem Lett 16, 401–426. https://doi.org/10.1007/s10311-017-0689-0
- 638 Gonzalez-Riano, C., Garcia, A., Barbas, C., 2016. Metabolomics studies in brain tissue: A review.
- Journal of Pharmaceutical and Biomedical Analysis, Review Issue 2016 130, 141–168.
- 640 https://doi.org/10.1016/j.jpba.2016.07.008
- Hamer, B., Hamer, D.P., Müller, W.E.G., Batel, R., 2004. Stress-70 proteins in marine mussel
 Mytilus galloprovincialis as biomarkers of environmental pollution: a field study.
- Environment International 30, 873–882. https://doi.org/10.1016/j.envint.2004.02.008
- Hassan, F., Nawaz, A., Rehman, M.S., Ali, M.A., Dilshad, S.M.R., Yang, C., 2019. Prospects of
- 645 HSP70 as a genetic marker for thermo-tolerance and immuno-modulation in animals under
- 646 climate change scenario. Animal Nutrition 5, 340–350.
 647 https://doi.org/10.1016/j.aninu.2019.06.005
- Huey, R.B., Slatkin, M., 1976. Cost and Benefits of Lizard Thermoregulation. The Quarterly
 Review of Biology 51, 363–384. https://doi.org/10.1086/409470
- Itoh, Y., Oishi, R., Nishibori, M., Saeki, K., 1986. Comparison of effects of phencyclidine and
 methamphetamine on body temperature in mice: a possible role for histamine neurons in
 thermoregulation. Naunyn-Schmiedeberg's Arch. Pharmacol. 332, 293–296.
 https://doi.org/10.1007/BF00504870
- Jin, J., Zhao, M., Wang, Y., Zhou, Z., Wan, F., Guo, J., 2020. Induced Thermotolerance and

655	Expression of Three Key Hsp Genes (Hsp70, Hsp21, and sHsp21) and Their Roles in the
000	Expression of three Key fisp Genes (htsp/6, fisp21, and stisp21) and then Koles in the
656	High Temperature Tolerance of Agasicles hygrophila. Frontiers in Physiology 10.
657	Joensuu, M., Wallis, T.P., Saber, S.H., Meunier, F.A., 2020. Phospholipases in neuronal function:
658	A role in learning and memory? Journal of Neurochemistry 153, 300-333.
659	https://doi.org/10.1111/jnc.14918
660	Kaur, M., Atif, F., Ansari, R.A., Ahmad, F., Raisuddin, S., 2011. The interactive effect of elevated
661	temperature on deltamethrin-induced biochemical stress responses in Channa punctata
662	Bloch. Chemico-Biological Interactions 193, 216–224.
663	https://doi.org/10.1016/j.cbi.2011.06.011
664	Kearney, M., Shine, R., Porter, W.P., 2009. The potential for behavioral thermoregulation to buffer
665	"cold-blooded" animals against climate warming. Proceedings of the National Academy of
666	Sciences 106, 3835–3840. https://doi.org/10.1073/pnas.0808913106
667	Kedderis, G.L., 2010. 1.07 - Biotransformation of Toxicants, in: McQueen, C.A. (Ed.),
668	Comprehensive Toxicology (Second Edition). Elsevier, Oxford, pp. 137-151.
669	https://doi.org/10.1016/B978-0-08-046884-6.00107-X
670	Kim, JK., Song, JY., Lee, JH., Park, DS., 2010. Physical characteristics and age structure of
671	Mongolian racerunner (Eremias argus; Larcertidae; Reptilia). J. Ecol. Environ. 33, 325-
672	331. https://doi.org/10.5141/JEFB.2010.33.4.325
673	Laitinen, P., Rämö, S., Nikunen, U., Jauhiainen, L., Siimes, K., Turtola, E., 2009. Glyphosate and
674	phosphorus leaching and residues in boreal sandy soil. Plant Soil 323, 267-283.
675	https://doi.org/10.1007/s11104-009-9935-y
676	Liu, W., Liu, P., Cui, L., Meng, Y., Tao, S., Han, X., Sun, B., 2022. Moderate climate warming

677	scenarios during embryonic and post-embryonic stages benefit a cold-climate lizard.
678	Functional Ecology 36, 1137–1150. https://doi.org/10.1111/1365-2435.14032
679	Liu, X., Shi, H., Liu, Z., Wang, J., Huang, J., 2019. Effect of Heat Stress on Heat Shock Protein 30
680	(Hsp30) mRNA Expression in Rainbow Trout (Oncorhynchus mykiss).
681	TrJFAS 19, 681–688.
682	Maggi, F., la Cecilia, D., Tang, F.H.M., McBratney, A., 2020. The global environmental hazard of
683	glyphosate use. Science of The Total Environment 717, 137167.
684	https://doi.org/10.1016/j.scitotenv.2020.137167
685	Pelosi, C., Bertrand, C., Bretagnolle, V., Coeurdassier, M., Delhomme, O., Deschamps, M., Gaba,
686	S., Millet, M., Nélieu, S., Fritsch, C., 2022. Glyphosate, AMPA and glufosinate in soils and
687	earthworms in a French arable landscape. Chemosphere 301, 134672.
688	https://doi.org/10.1016/j.chemosphere.2022.134672
689	Pochron, S., Choudhury, M., Gomez, R., Hussaini, S., Illuzzi, K., Mann, M., Mezic, M., Nikakis,
690	J., Tucker, C., 2019. Temperature and body mass drive earthworm (Eisenia fetida)
691	sensitivity to a popular glyphosate-based herbicide. Applied Soil Ecology 139, 32-39.
692	https://doi.org/10.1016/j.apsoil.2019.03.015
693	Price, E.R., Sirsat, T.S., Sirsat, S.K.G., Kang, G., Keereetaweep, J., Aziz, M., Chapman, K.D.,
694	Dzialowski, E.M., 2017. Thermal acclimation in American alligators: Effects of
695	temperature regime on growth rate, mitochondrial function, and membrane composition.
696	Journal of Thermal Biology, SI:Ectotherms: Performance 68, 45-54.
697	https://doi.org/10.1016/j.jtherbio.2016.06.016
698	Rich, C.N., Talent, L.G., 2009. SOIL INGESTION MAY BE AN IMPORTANT ROUTE FOR

- 700 311. https://doi.org/10.1897/08-035.1
- 701 Rourke, J.L., Sinal, C.J., 2014. Biotransformation/Metabolism, in: Wexler, P. (Ed.), Encyclopedia
- 702 of Toxicology (Third Edition). Academic Press, Oxford, pp. 490–502.
 703 https://doi.org/10.1016/B978-0-12-386454-3.00007-5
- 704 Rutschmann, A., Dupoué, A., Miles, D.B., Megía-Palma, R., Lauden, C., Richard, M., Badiane, A.,
- 705 Rozen-Rechels, D., Brevet, M., Blaimont, P., Meylan, S., Clobert, J., Le Galliard, J., 2021.
- 706 Intense nocturnal warming alters growth strategies, colouration and parasite load in a
- 707 diurnal lizard. Journal of Animal Ecology 90, 1864–1877. https://doi.org/10.1111/1365-
- 708 2656.13502
- 709 Seebacher, F., 2005. A review of thermoregulation and physiological performance in reptiles: what
- is the role of phenotypic flexibility? J Comp Physiol B 175, 453–461.
 https://doi.org/10.1007/s00360-005-0010-6
- Seebacher, F., Murray, S.A., 2007. Transient Receptor Potential Ion Channels Control
 Thermoregulatory Behaviour in Reptiles. PLOS ONE 2, e281.
 https://doi.org/10.1371/journal.pone.0000281
- Seebacher, F., White, C.R., Franklin, C.E., 2015. Physiological plasticity increases resilience of
 ectothermic animals to climate change. Nature Clim Change 5, 61–66.
 https://doi.org/10.1038/nclimate2457
- Silva, L.C.M., Daam, M.A., Gusmao, F., 2020. Acclimation alters glyphosate temperaturedependent toxicity: Implications for risk assessment under climate change. Journal of
 Hazardous Materials 385, 121512. https://doi.org/10.1016/j.jhazmat.2019.121512

721	Sinervo, B., Méndez-de-la-Cruz, F., Miles, D.B., Heulin, B., Bastiaans, E., Villagrán-Santa Cruz,
722	M., Lara-Resendiz, R., Martínez-Méndez, N., Calderón-Espinosa, M.L., Meza-Lázaro,
723	R.N., Gadsden, H., Avila, L.J., Morando, M., De la Riva, I.J., Sepulveda, P.V., Rocha,
724	C.F.D., Ibargüengoytía, N., Puntriano, C.A., Massot, M., Lepetz, V., Oksanen, T.A.,
725	Chapple, D.G., Bauer, A.M., Branch, W.R., Clobert, J., Sites, J.W., 2010a. Erosion of
726	Lizard Diversity by Climate Change and Altered Thermal Niches. Science 328, 894-899.
727	https://doi.org/10.1126/science.1184695
728	Sinervo, B., Méndez-de-la-Cruz, F., Miles, D.B., Heulin, B., Bastiaans, E., Villagrán-Santa Cruz,
729	M., Lara-Resendiz, R., Martínez-Méndez, N., Calderón-Espinosa, M.L., Meza-Lázaro,
730	R.N., Gadsden, H., Avila, L.J., Morando, M., De la Riva, I.J., Sepulveda, P.V., Rocha,
731	C.F.D., Ibargüengoytía, N., Puntriano, C.A., Massot, M., Lepetz, V., Oksanen, T.A.,
732	Chapple, D.G., Bauer, A.M., Branch, W.R., Clobert, J., Sites, J.W., 2010b. Erosion of
733	Lizard Diversity by Climate Change and Altered Thermal Niches. Science 328, 894-899.
734	https://doi.org/10.1126/science.1184695
735	Spellerberg, I.F., 1972. Thermal ecology of allopatric lizards (Sphenomorphus) in Southeast
736	Australia. Oecologia 11, 1–16. https://doi.org/10.1007/BF00345706
737	Stocker, T.F., D., Q., 2013. IPCC, 2013: Climate Change 2013: The Physical Science Basis.
738	Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental
739	Panel on Climate Change. Cambridge University Press, Cambridge, United Kingdom and
740	New York, NY, USA, 1535 pp.
741	Sulukan, E., Baran, A., Şenol, O., Kankaynar, M., Yıldırım, S., Bolat, İ., Ceyhun, H.A., Toraman,
742	E., Ceyhun, S.B., 2023. Global warming and glyphosate toxicity (I): Adult zebrafish

743	modelling with behavioural, immunohistochemical and metabolomic approaches. Science
744	of The Total Environment 858, 160086. https://doi.org/10.1016/j.scitotenv.2022.160086
745	Sviridov, A.V., Shushkova, T.V., Ermakova, I.T., Ivanova, E.V., Epiktetov, D.O., Leontievsky,
746	A.A., 2015. Microbial degradation of glyphosate herbicides (Review). Appl Biochem
747	Microbiol 51, 188–195. https://doi.org/10.1134/S0003683815020209
748	Todd, B., Willson, J., Gibbons, J., 2010. The Global Status of Reptiles and Causes of Their Decline,
749	in: Sparling, D., Linder, G., Bishop, C., Krest, S. (Eds.), Ecotoxicology of Amphibians and
750	Reptiles, Second Edition. CRC Press, pp. 47–67. https://doi.org/10.1201/EBK1420064162-
751	c3
752	Vijay K. Nandula, 2010. Glyphosate resistance in crops and weeds: history, development, and
753	management. John Wiley & Sons, 2010.
754	Wada, H., Inagaki, N., Itowi, N., Yamatodani, A., 1991. Histaminergic neuron system in the brain:
755	Distribution and possible functions. Brain Research Bulletin 27, 367-370.
756	https://doi.org/10.1016/0361-9230(91)90126-5
757	Wang, Z., Liu, Ran, Zhang, L., Yu, S., Nie, Y., Deng, Y., Liu, Rui, Zhu, W., Zhou, Z., Diao, J.,
758	2022. Thermoregulation of Eremias argus alters temperature-dependent toxicity of beta-
759	cyfluthrin: Ecotoxicological effects considering ectotherm behavior traits. Environmental
760	Pollution 293, 118461. https://doi.org/10.1016/j.envpol.2021.118461
761	Wang, Z., Zhu, W., Xu, Y., Yu, S., Zhang, L., Zhou, Z., Diao, J., 2021. Effects of simazine and food
762	deprivation chronic stress on energy allocation among the costly physiological processes
763	of male lizards (Eremias argus). Environmental Pollution 269, 116139.
764	https://doi.org/10.1016/j.envpol.2020.116139

765	Wee, J., Lee, YS., Kim, Y., Son, J., Cho, K., 2021. Temperature and Aging Affect Glyphosate
766	Toxicity and Fatty Acid Composition in Allonychiurus kimi (Lee) (Collembola). Toxics 9,
767	126. https://doi.org/10.3390/toxics9060126
768	Withers, P.C., Campbell, J.D., 1985. Effects of Environmental Cost on Thermoregulation in the
769	Desert Iguana. Physiological Zoology 58, 329–339.
770	https://doi.org/10.1086/physzool.58.3.30156004
771	Wu, W., Lin, Y., 2020. Determination of glyphosate in soils by HPLC with pre-column
772	derivatization using 4-chloro-7-nitro-1,2,3-benzoxadiazole. Chinese Journal of Pesticide
773	Science 22(6), 1027–1032. https://doi.org/10.16801/j.issn.1008-7303.2020.0154
774	Xiao, R., Xu, X.Z.S., 2021. Temperature Sensation: From Molecular Thermosensors to Neural
775	Circuits and Coding Principles. Annu. Rev. Physiol. 83, 205–230.
776	https://doi.org/10.1146/annurev-physiol-031220-095215
777	Yu, H., Zhang, C., Xing, W., Li, T., Xu, G., Ma, Z., Jiang, N., Luo, L., 2020. Comparative study on
778	the non-specific immune response and hsp70 gene expression among three strains of koi
779	(Cyprinus carpio) under acute heat stress. Aquaculture Reports 18, 100461.
780	https://doi.org/10.1016/j.aqrep.2020.100461
781	Zabaloy, M.C., Allegrini, M., Hernandez Guijarro, K., Behrends Kraemer, F., Morrás, H., Erijman,
782	L., 2022. Microbiomes and glyphosate biodegradation in edaphic and aquatic environments:
783	recent issues and trends. World J Microbiol Biotechnol 38, 1-29.
784	https://doi.org/10.1007/s11274-022-03281-w
785	Zhang, Y., Guo, Q., Tan, D., He, Z., Wang, Y., Liu, X., 2018. Effects of low-levels of three
786	hexabromocyclododecane diastereomers on the metabolic profiles of pak choi leaves using

- 787 high-throughput untargeted metabolomics approach. Environmental Pollution 242, 1961–
- 788 1969. https://doi.org/10.1016/j.envpol.2018.07.062
- 789 Zhao, E.M., Zhao, K.T., Zhao, K.Y., 1999. Reptilia: Vol. 2. Squamata: Lacertilia, in: Fauna Sinica.
- 790 Beijing Science Press, Beijing.
- 791